

PAPER**PATHOLOGY/BIOLOGY**

Mark Eric Benbow,¹ Ph.D.; Jennifer L. Pechal,¹ Ph.D.; Jennifer M. Lang,² B.S.; Racheal Erb,³ B.S.; and John R. Wallace,³ Ph.D.

The Potential of High-throughput Metagenomic Sequencing of Aquatic Bacterial Communities to Estimate the Postmortem Submersion Interval*

ABSTRACT: Human remains can be discovered in freshwater or marine ecosystems, circumstances where insects and other invertebrates have infrequently been used for understanding the time of postmortem submersion. In this study, the identification and succession of epinecrotic bacterial communities on vertebrate remains were described during decomposition in a temperate headwater stream during two seasons (summer and winter). Bacterial communities were characterized with 454 pyrosequencing and analyzed at phyletic and generic taxonomic resolutions. There was a significant increase in genera richness over decomposition during both seasons. Additionally, multivariate statistical modeling revealed significant differences in bacterial communities between seasons at both taxonomic resolutions and significant genera differences among sampling days within each season, suggesting a succession of these communities. These data are the first to describe aquatic bacterial succession using high-throughput metagenomic sequencing on vertebrate remains submerged in a freshwater habitat, and provide initial evidence for their potential use in forensic investigations.

KEYWORDS: forensic science, aquatic microbial communities, epinecrotic communities, high-throughput metagenomic sequencing, pyrosequencing, postmortem submersion interval, necrobiome

Over the last several years, there has been an increasing recognition of the importance of establishing a strong basic science foundation in the forensic sciences (1,2). With the development, reliability, and increasingly widespread use of high-throughput metagenomic sequencing (e.g., 454 pyrosequencing and Illumina MiSeq), the forensic sciences are poised to harness “-omics” technology to bridge basic science with medicolegal applications. This unique application of technology is especially timely for those forensic disciplines without traditional genomic foundations, such as taphonomy. This application of microbial communities also has the potential to expand into areas of forensic investigations involving human remains discovered outdoors, much like recent studies that have used high-throughput metagenomic sequencing of a human (3), swine (4), and mouse (5) microbiomes. The use of these next-generation technologies has the *potential* to enhance

the ability of investigators to provide more accurate estimates and provide error rates associated with evidentiary interpretation, especially in quantitative timelines related to death.

Recently, metagenomic tools have become more widely available, thus allowing thorough investigations into epinecrotic bacterial communities during vertebrate decomposition (3–5). Pyrosequencing and other high-throughput metagenomic approaches (e.g., Illumina MiSeq) can taxonomically identify bacterial community changes throughout decomposition, generate large amounts of data, characterize sequences of unculturable or newly discovered microbial species, and are established genomic methods (6–8). Analyses using metagenomics provide the depth of results required to more comprehensively characterize and understand temporal and spatial bacterial community composition and succession on decomposing vertebrate remains (4). Given the technological advancements and increase in computing power and statistical modeling capabilities for metagenomic data, the characterization of environmental microbial communities using metagenomics is now becoming more widely available and cost-effective. The forensic sciences are positioned well to take advantage of the novel and innovative assessments and statistical modeling approaches for using these ubiquitous biological communities as evidence.

One task for medicolegal forensic investigators is the scientific estimation of the amount of time human remains have been submerged or partially submerged in aquatic environmental conditions, such as rivers, lakes, marshes, or the ocean, in cases involving suspicious drowning or aquatic body disposal. This

¹Department of Entomology, Michigan State University, East Lansing, MI 48824.

²Department of Biology, University of Dayton, Dayton, OH 45469-2320.

³Department of Biology, Millersville University, Millersville, PA 17551.

*Funding for this project was provided by the small grants program of the Pathology/Biology Section of the American Academy of Forensic Sciences and through discretionary funds of the Department of Biology at the University of Dayton. Additionally, this work has been supported in part by the University of Dayton Office of Graduate Professional & Continuing Education through the Graduate Student Summer Fellowship Program awarded to JML and the Millersville University Student Research Grant awarded to RE.

Received 23 April 2014; and in revised form 10 Nov. 2014; accepted 14 Nov. 2014.

time frame has been defined as the postmortem submersion interval (PMSI): the time between when a corpse enters an aquatic environment until discovery by authorities (9–11). The aquatic decomposition process of vertebrate remains has been described as initiating with submerged fresh remains proceeding to floating decay stages and ending with sunken remains (9–11). However, these broad categorizations and visual assessments of decomposition can vary based on water quality, depth of the remains, and other abiotic and biotic influences (9). Therefore, other biological indicators, such as microbial communities that are ubiquitous within environment, have the potential to better estimate a PMSI range.

In aquatic ecosystems, microbes are predominately found within biofilms rather than as free-floating planktonic forms (12,13). Biofilms are matrix-enclosed microbial communities that are both trophically (heterotrophs and autotrophs) and phylogenetically (prokaryotes and eukaryotes) diverse, containing algal, bacterial, fungal, and protozoan species (14,15). These microorganisms are encased in an extracellular polymeric substance (EPS) that is attached to surfaces in aqueous solutions (12,16). The EPS matrix is protective against changing abiotic conditions (17,18), traps and stores nutrients (14,19,20), and accrues enzymes that break down organic matter (21,22). Essentially, the biofilm is a micro-ecosystem that can be described using ecological theory and principles of landscape ecology (15). The type of surface classifies the biofilm and influences community composition and energy dynamics by dictating the dominant community. For example, epilithic biofilms that form on inorganic substrates like rocks are more autotrophic and represented by an abundant algal community, while the epixylic biofilms found on decomposing plant matter are more heterotrophic with a substantial fungal community (23–25). Within biofilms studied on decomposing piglet carcasses, diatoms (unicellular algae with a silica cell wall) demonstrated patterns of diversity and abundance throughout decomposition (9,26), but identification of these organisms is taxonomically intensive and requires habitat conditions with abundant sunlight. Bacteria diversity and abundance also change during marine decomposition, but cloning methods cannot effectively identify the entire community taxa composition (27). Biofilms on decomposing vertebrate carcasses, recently classified as epinecrotic (4), have been noted on salmon (28), waterfowl (29), rats (30), and pigs (11), but only a few studies have focused on this biofilm community for use in forensics. The commonality of these studies is that they are based on tracking community composition changes as epinecrotic biofilms develop during the decomposition process of a carrion substrate.

Biofilm development typically follows a pattern of succession, or the natural sequence of species change and replacement over time. Bacteria are the first colonizers and establish the pioneer community (31); these initial colonizers remain as the base layer for subsequent immigration and establishment of larger microorganisms (14,15). As the biofilm ages, larger and filamentous algae and fungi are incorporated until filamentous organisms dominate the biofilm at a mature stage with composition depending on the substrate and light conditions. Fungal organisms could play a role in the use of epinecrotic communities for forensics, but they have yet to be studied using high-throughput metagenomics. It is important to note that bacteria are always present throughout succession and are an important biological indicator of biofilm succession (27). Succession can range from weeks to months depending on abiotic and biotic conditions, and this sequence of bacterial taxa replacement during vertebrate decomposition could potentially be used to estimate the physiological

time (e.g., accumulated degree days) in aquatic habitats, as it has been demonstrated in a terrestrial habitat (4). There are several factors known to affect microbial community succession during vertebrate carrion decomposition in terrestrial habitats including temperature (32,33), moisture and humidity (34), tissue type (27), surrounding vegetation (35,36), and soil pH (37). Likewise, there are a variety of environmental factors (e.g., temperature, dissolved oxygen, carbon, and nutrients) that researchers studying microbial assemblages of aquatic systems have identified as important drivers of microbial community succession (38).

In terrestrial environments, insects, with well-known life-history traits (e.g., development rates), along with geographical and seasonal distribution data, are often used to assist in minimum postmortem interval (mPMI) estimates, but only recently have microbial communities been investigated for similar applications (3–5). For submerged remains in aquatic habitats, literary consensus is that very few, if any, invertebrate species have evolved to consume vertebrate remains in a way that can reliably be used for estimating a PMSI range (10). However, several empirical studies using traditional sequencing approaches have provided valuable baseline information on the potential usefulness of using naturally occurring diverse assembly of microbes, such as algae, in aquatic systems to form evidentiary linkages among suspects, victims, and physical evidence (e.g., weapons) (9,11,39–44). Indeed, Keiper and Casamatta (45) suggested additional studies of algae and other aquatic flora for use in forensics. In response, two studies were conducted and each reported that algal and diatom communities showed promise for being used in estimating a PMSI (9,11), but both studies did not evaluate the prokaryotic taxa that are also part of aquatic biofilm communities. However, Dickson et al. (27) documented the bacteria community succession (but not the fungal communities) on decomposing remains in marine habitats using swine heads as a model of decomposition and employing capillary electrophoresis sequencing methods. While bacterial assemblages are ubiquitous, there have been no studies to date that have used metagenomic sequencing similar to what has been done in terrestrial settings (3,4). Thus, a description of the prokaryotic component of the microbial community throughout the decomposition process could be important in estimating a PMSI range of human remains in aquatic environments.

In this study, we employed high-throughput metagenomic sequencing to describe bacterial communities on the external surfaces of swine remains [as surrogates for human corpses (46)] to better understand the potential use of epinecrotic bacterial succession in aquatic death investigations. The objectives of this research were threefold: (i) characterize how the epinecrotic bacteria communities of decomposing vertebrate carcasses changed during decomposition in a stream habitat, (ii) assess whether this succession was consistent between two seasons (summer vs. winter), and (iii) provide data important for discussion of the potentials and limitations for using aquatic epinecrotic community succession in future studies of forensic application.

Methods

Study Site Description

The study was conducted from 26 June to 17 July 2012 (summer trial), and 9 November to 21 December 2012 (winter trial), in a first-order tributary to the west branch of Big Spring Run (39°59'29.1"N 76°15'49.0"W) within the Conestoga River watershed of Lancaster, Pennsylvania. Despite flowing through a

suburban/agricultural area of the watershed, the stream was bordered by 10–20 m of riparian forest and shrub vegetation. Canopy trees within the riparian zone that shaded 90% of the stream channel included silver maple (*Acer saccharinum* L.), box elder (*Acer negundo* L.), and sycamore (*Platanus occidentalis* L.) with a ground cover of multiflora rose (*Rosa multiflora* Thunb.) bordering both stream banks. The stream substrate consisted of a mixture of pebble and cobble throughout the study site. The swine carcasses (see below for details) were individually placed in run habitats of the stream to avoid being silted over in pools and the abrasive action of riffle habitats.

Water Chemistry

Water quality parameters of dissolved oxygen (mg/L), pH, conductivity (mS/cm), water temperature (°C), total dissolved solids (g/L), oxidation–reduction potential (mV), and salinity (ppt) were measured at a single location 30 m upstream of the uppermost carcass and 30 m downstream of the lowermost carcass on each sampling day during both seasons using a Horiba® (Kyoto, Japan) Multi Water Quality Checker (U-50 Series).

Qualitative Decomposition Changes

To visually describe the decomposition process for the carcasses, the duration of each decay stage was assessed and described using the framework of Zimmerman and Wallace (9). The summer trial ended the day when the sunken remains stage was sampled as there was minimal skin remaining for swabbing for epinecrotic biofilm, while the winter trial ended unexpectedly and abruptly as a result of scavenging events and cage vandalism from the public on Day 42 of decomposition. See below for more information regarding the sampling times.

Epinecrotic Bacterial Community Sample Collection

Stillborn *Sus scrofa* L. carcasses (1–1.5 kg; $n = 3$ per trial) obtained from the Penn State University Swine Research Facility (State College, PA) were used as the decomposition model for epinecrotic biofilms formation during summer and winter trials. Swine carrion has been evaluated as a suitable surrogate for human cadavers in decomposition studies (46), and swine skin has been used as a model for laboratory human skin bacteriological studies (47–50); however, piglet carcasses were used in this study and may not represent the same decay process of adult swine but do provide good models for research. Carcasses were placed on plastic drawer organizer trays ($0.38 \times 0.15 \times 0.05$ m) inside small, metal game traps ($0.61 \times 0.18 \times 0.18$ m, Havahart®; Animals B-Gone, Orrstown, PA) to facilitate sampling as the carcass disarticulated and to prevent scavenger removal of the carcasses. The cages also prevented the carcasses from breaking the surface, subsequently inhibiting colonization by terrestrial insects. Traps were anchored to the streambed on previously secured rebar, which allowed easy trap removal for sampling. During each trial, carcasses were placed ~15–20 m downstream from each other with one riffle/run reach between each one along the study segment of stream to minimize upstream effects of any carcass. This flowing nature of streams makes it difficult to have completely independent carcasses, and we acknowledge this as a limitation to this study.

Epinecrotic biofilm was sampled weekly at four consecutive time points during the summer trial [29 June (Day 0), 3 July (Day 7), 10 July (Day 14), 17 July (Day 21)] and biweekly

during the winter trial [9 November (Day 0), 23 November (Day 14), 7 December (Day 28), 21 December (Day 42)] to account for the delayed decomposition resulting from cooler temperatures. To characterize the communities prior to submersion, samples were collected on “Day 0” prior to the carcasses being placed into the stream and thus represent the initial bacterial communities of the carcasses. The bacterial samples were collected from the external surface of the abdomen/rib cage using individually packaged sterile swabs; an area (approximately 5×1 cm) was swabbed with six strokes where the swab was rotated 180° after the third stroke and one direction counted as a stroke with care taken to not duplicate any previously sampled areas. The swab sample was individually placed into a sterile microcentrifuge tube and stored on ice for approximately 0.5 h until samples were placed in a –20°C freezer for further DNA extractions. The most downstream carcass was sampled first and sampling progressed upstream on each date. Carcasses were immediately submerged upon sample collection completion, and new gloves were worn for each sample collection.

DNA Extraction

DNA was extracted using a combination of methods from Miller et al. (51) and Zhou et al. (52) as suggested by Lear et al. (53). Samples were lysed in 1 mL extraction buffer made of 100 mM Tris–HCl (pH 8.0), 100 mM EDTA disodium salt (pH 8.0), 100 mM sodium phosphate (pH 8.0), 1.5 M sodium chloride, and 1% CTAB; 20 µL of proteinase K (10 mg/mL) and 25 µL of SDS (20%) were added to each sample in the extraction buffer prior to bead beating (0.25 g each of 0.1 mm and 0.5 mm glass beads) for 15 min on a horizontal vortex adaptor (MO BIO Laboratories, Carlsbad, CA) at full speed. The samples were incubated at 60°C for 30 min with gentle end-over-end inversions by hand after 15 min; 250 µL of supernatant was collected in a new microcentrifuge tube after centrifugation at 6,000 g for 10 min. The lysis process was repeated without vortexing with an additional 500 µL extraction buffer, 10 µL proteinase K, and 12.5 µL SDS to obtain a final supernatant volume of approximately 750 µL. DNA was separated from organic debris with a chloroformisoamyl alcohol (24:1 vol/vol) extraction and precipitated overnight at –20°C using isopropanol. Samples were removed from the –20°C and warmed to 37°C to dissolve salt precipitates, and the DNA was pelleted at 15,000 g for 30 min. Finally, the DNA pellet was washed twice with ice-cold 70% ethanol and dissolved in 50–100 µL ultrapure water (NANOpure II™, Thermo Scientific, Waltham, MA) water, depending on the DNA pellet size.

454-Pyrosequencing

Epinecrotic bacterial community composition was characterized by modified bacterial tagged encoded FLX amplicon pyrosequencing (bTEFAP) as described by Pechal et al. (4) for the characterization of bacterial communities from swine remains in terrestrial habitats. Briefly, extracted DNA (>200 ng) was sent to the Molecular Research Laboratory (Shallowater, TX). PCR amplification for bacterial V1-3 regions of 16S rRNA using Gray28F (5′TTTGATCCTGGCTCAG) and Gray519r (5′GTNTTACNGCGGCKGCTG) primers was carried out, as previously described (54–58). Sequencing library generation began with a one-step PCR using a mixture of Hot Start and HotStar high-fidelity *Taq* polymerases. Following RTL protocols (www.researchandtesting.com), a Roche 454 FLX instrument

with titanium reagents and titanium procedures was used to perform the tag-encoded FLX amplicon pyrosequencing analyses. The sequences were deposited in the Sequence Read Archive (SRA) at the European Bioinformatics Institute with study accession numbers PRJEB5310 and PRJEB5311.

454-Pyrosequencing Analyses

The Q25 sequence data derived from the sequencing process was processed using a proprietary analysis pipeline (www.mrd-nalab.com). Sequences were depleted of barcodes and primers; then, any sequences with <200 bp, ambiguous base calls, or with homopolymer runs exceeding 6 bp were removed. Finally, sequences were denoised and chimeras removed. Operational taxonomic units (OTUs) were defined after the removal of singleton sequences and with clustering at 3% divergence (59–63). OTUs were taxonomically classified using BLASTn against a curated GreenGenes database (64). All three carcasses during the summer trial had successful sequencing results, but unfortunately during the winter trial, there were insufficient reads for one of the carcasses and thus that individual was excluded from subsequent data analyses.

Statistical Analyses

To test for water chemistry difference upstream (US) and downstream (DS) of the carcasses within each seasonal trial, a paired t-test was employed using Prism 5 (GraphPad Software, Inc., La Jolla, CA) by the measurements taken on each sampling date and the pairing date. To assess bacterial community structure change over decomposition, the taxon richness was calculated according to Simpson (65) and mean relative taxon abundance was determined at both phyletic and generic taxonomic resolutions for individual replicate carcasses and also averaged for each date. Further, Bray–Curtis distance with nonmetric multidimensional scaling (NMDS) and 999 permutations using vegan 2.0-7 library in the R statistical package (66) was used to visualize bacterial community differences between seasons and among sample weeks. For each ordination, the two axes that explained the most variation and the strongest

orthogonality (lowest stress) were used for representing the data in multidimensional space (67). Differences in bacterial communities and variation over decomposition (i.e., sampling dates) were tested using Bray–Curtis dissimilarity values with permutational multivariate analysis of variance (PERMANOVA) with the adonis function using the vegan library in the R statistical package (66). PERMANOVA is a nonparametric technique used to differentiate groups of data based on a dissimilarity matrix (68). Significance was determined at $\alpha = 0.05$ for all analyses.

Results

Water Chemistry

Water temperature, pH, and total dissolved solids (TDS) were generally greater during the summer trial (Table 1), while dissolved oxygen (DO) and oxidation–reduction potential (ORP) were higher in the winter trial. The other variables showed consistency between seasons. Within each seasonal trial, the only water quality variable that was significantly different was DO ($t = 3.38$, $df = 3$, $p = 0.043$), which was higher upstream (US) than downstream (DS) of the carcasses only during the summer trial.

Qualitative Decomposition Changes

For the summer trial, the mean (\pm standard deviation) number of days to reach the first day of each decomposition stage was the following: submerged fresh began at 0.0 (± 0.0) days, early floating began at 5.7 (± 2.3) days, early floating decay at 9.0 (± 1.7) days, advanced floating decay at 14.0 (± 0.0) days, and sunken remains at 22.0 (± 1.7) days (Table S1). During the winter trial, there was a delayed onset for each decomposition stage with submerged fresh, early floating, and early floating decay occurring at days 0.0 (± 0.0), 5.0 (± 5.6), and 21.0 (± 0.0), respectively (Table S1); unfortunately, there was an unforeseen scavenging event and vandalism that caused the abrupt end of this trial and prevented further documentation of the decomposition progression.

TABLE 1—Water quality parameters were collected at a single location 30 m upstream (US) and downstream (DS) of the swine carcasses on each sampling day for each trial. Paired t-tests were used to evaluate statistical differences between US and DS locations for each water chemistry parameter. DO was the only parameter that was found to be significantly different between US and DS: DO was higher US than DS only during the summer trial.

Season	Day	Sampling Location	DO (mg/L)	pH	SC (mS/cm)	Temp (°C)	TDS (g/L)	ORP (mV)	Salinity (ppt)
Summer	Day 0	US	4.46	8.13	0.936	20.7	0.468	65.6	0.46
		DS	3.54	8.07	0.920	20.2	0.460	63.3	0.45
	Day 7	US	7.22	8.73	1.043	20.7	0.522	68.9	0.52
		DS	6.38	8.20	1.032	19.9	0.516	64.7	0.51
	Day 14	US	5.24	8.13	0.973	21.8	0.487	38.4	0.48
		DS	2.86	8.11	0.842	22.6	0.421	36.0	0.41
	Day 21	US	6.23	8.12	0.967	23.2	0.438	36.5	0.48
		DS	3.32	8.15	0.958	23.8	0.479	36.8	0.47
Winter	Day 0	US	3.23	—	1.010	10.6	0.413	95.6	0.41
		DS	3.12	—	0.878	10.2	0.439	711	0.44
	Day 14	US	10.51	8.24	0.941	13.6	0.603	207	0.50
		DS	9.72	8.21	0.930	11.6	0.596	196	0.50
	Day 28	US	9.44	7.79	0.976	11.7	0.626	242	0.50
		DS	10.17	8.01	0.972	10.3	0.624	258	0.50
	Day 42	US	9.26	7.93	0.498	9.6	0.321	250	0.20
		DS	9.5	8.05	0.820	9.6	0.525	257	0.40

DO, dissolved oxygen; SC, specific conductivity; TDS, total dissolved solids; ORP, oxidation–reduction potential.

The mean (\pm standard deviation) duration (number of days) the carcasses were in each decomposition stage during the summer trial was as follows: submerged fresh for 5.7 (\pm 2.3) days, early floating for 3.3 (\pm 0.6) days, early floating decay for 5.0 (\pm 1.7) days, advanced floating decay for 8.0 (\pm 1.7) days, and sunken remains for 3.3 (\pm 0.6) days (Table S1). The sunken remains stage may have lasted longer, but our sampling was completed after a single collection of the sunken remains. Decomposition progression was slower during the winter trial with the replicate carcasses characterized as submerged fresh for 7.5 (\pm 5.0) days, early floating for 16.0 (\pm 5.6) days, and early floating decay for 28.0 (\pm 0.0) days (Table S1); again, the unforeseen scavenging event and vandalism abruptly ended this trial.

Epinecrotic Bacterial Communities

The epinecrotic bacterial communities of the decomposing swine carcasses in the stream were represented by 17 phyla and 179 genera in the summer trial, while 13 phyla and 202 genera were documented in the winter trial (Tables S2 and S3). There were significant changes in taxon richness over decomposition with peak mean richness increasing from 21 to 82 genera within the first 21 days of decomposition during the summer, and peak mean richness increasing from 38 to 109 genera after 42 days of decomposition in the winter. Proteobacteria and Firmicutes were the predominate phyla during both summer and winter trials. However, the winter carcasses never reached the advanced floating remains or sunken remains stages, making it difficult to compare community differences seasons by stage. Over the summer trial, the mean relative abundance of Proteobacteria decreased from 91.4, 74.5, 60.4 to 30.5% over the four continuous weekly sampling periods, while Firmicutes displayed an inverse trend of increasing mean relative abundance from 2.4,

23.6, 36.3 to 58.0% over the same time period (Fig. 1). The relative abundance of the epinecrotic taxa for each individual carcass also demonstrated a similar inverse relationship, with Proteobacteria decreasing as Firmicutes increased over the decomposition process during both summer (Fig. 2) and winter (Fig. 3) trials.

At the generic level during the summer trial, the initial sampling period (Day 0) was dominated by *Pseudomonas* (42.3%), *Psychrobacter* (36.1%), and *Ewingella* (11.1%); by Day 7, *Klebsiella* (13.0%), *Psychrobacter* (12.5%), and *Citrobacter* (9.4%) were the predominate genera; on Day 14, there was a shift in community structure with *Zoogloea* (17.2%), *Clostridium* (15.0%), *Dechloromonas* (14.2%), and *Desulfosporomusa* (8.2%) being the most abundant genera; and finally on Day 21, *Clostridium* (25.2%), *Enterococcus* (16.9%), and *Lactobacillus* (9.3%) were the predominant genera (Table S2).

A similar succession pattern was documented during the winter trial, as Proteobacteria mean relative abundance decreased from the initial sampling on Day 0 (81.9%) to Day 14 (52.3%) to Day 28 (36.7%), but was consistent on Day 42 (38.2%) of sampling (Fig. 1). Firmicutes mean relative abundance increased from the initial sampling on Day 0 (14.1%) to Day 14 (47.0%) to Day 28 (61.7%), but decreased slightly on Day 42 (56.5%) of sampling (Fig. 1).

At the generic level during the winter trial, the initial sampling time (Day 0) was dominated by *Psychrobacter* (41.7%), *Pseudomonas* (31.5%), and *Enterococcus* (9.1%); *Lactococcus* (27.8%) was predominant on Day 14; Day 28 was dominated by *Proteocatella* (30.3%), *Clostridium* (10.4%), and *Veillonellaceae* (10.2%); and finally on Day 42, *Veillonellaceae* (18.6%) and *Clostridium* (12.4%) were still the predominant genera (Table S3).

Using NMDS (Figs 4 and 5) and PERMANOVA (Table 2), the epinecrotic bacterial communities were significantly different

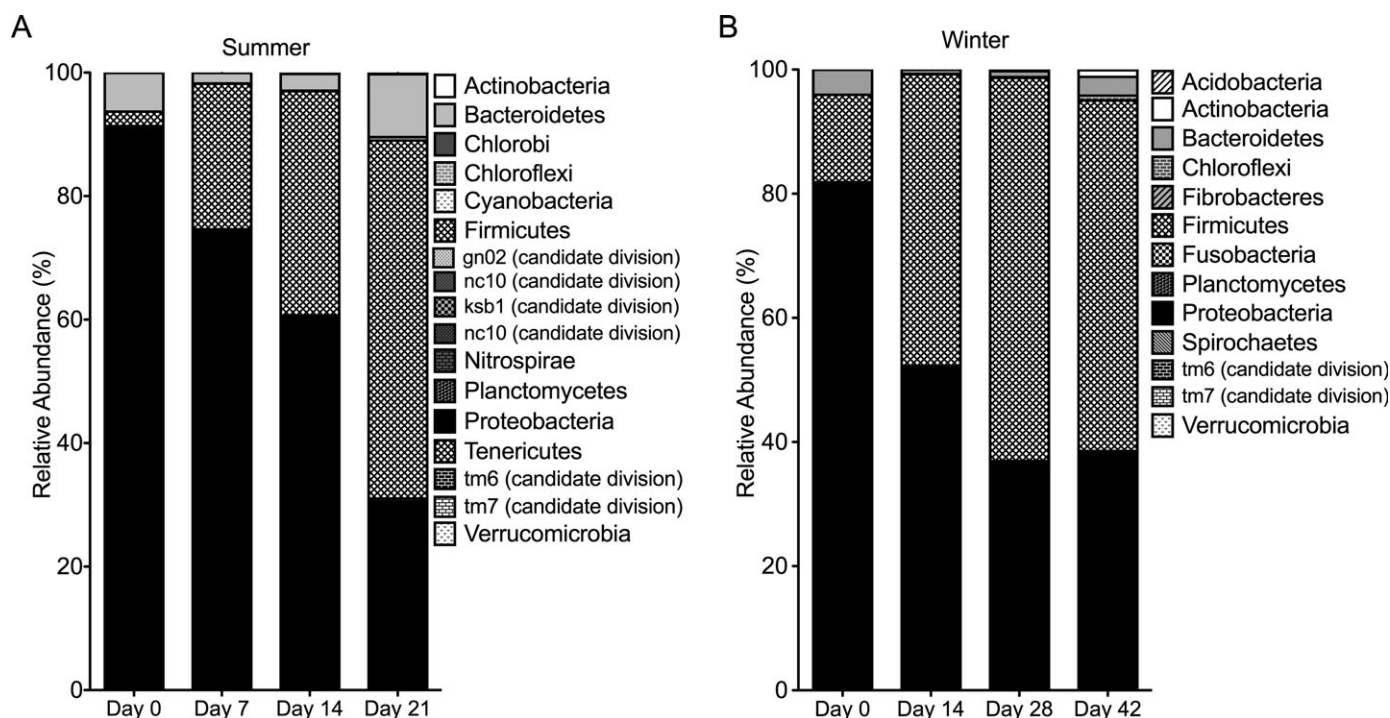


FIG. 1—Epinecrotic bacterial community taxa mean relative abundances over decomposition of swine carcasses submerged in a stream: phyla are presented for summer (A) and winter (B) trials. Genera relative abundances are provided in Tables S1 and S2 for the summer and winter trials, respectively.

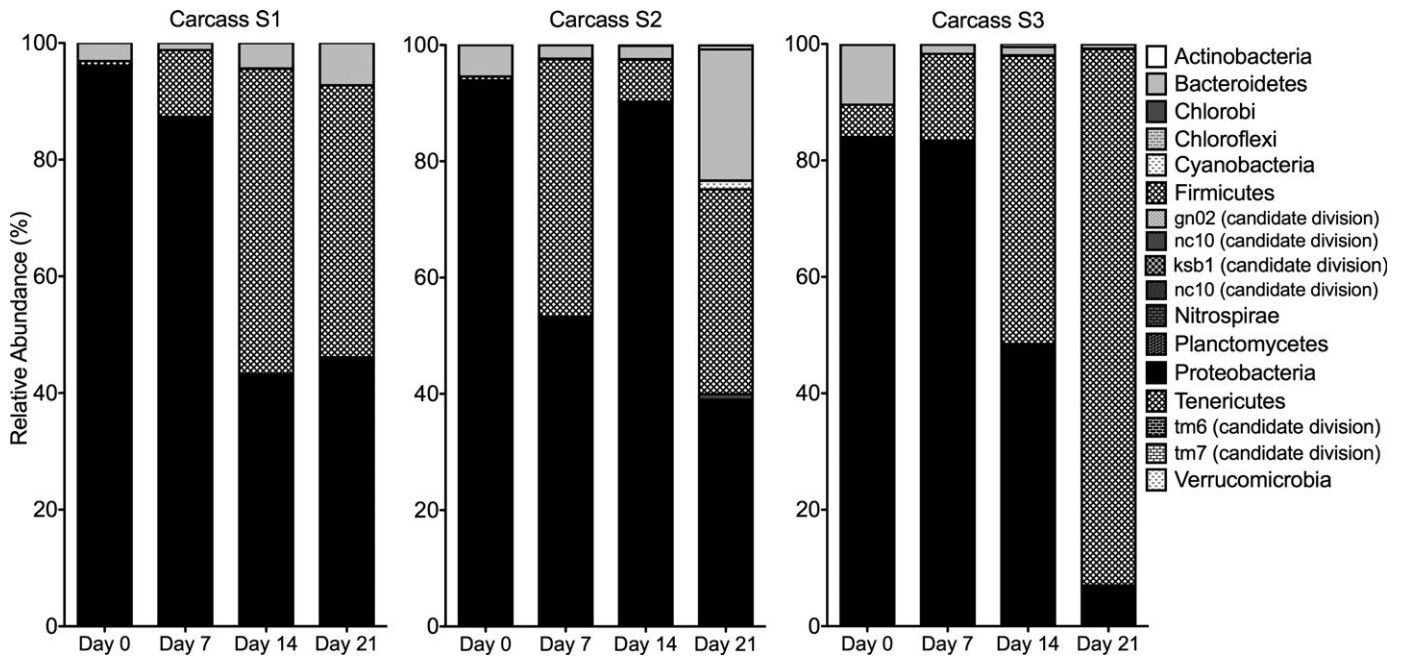


FIG. 2—Epinecrotic bacterial community taxa relative abundances change over decomposition of individual swine carcasses submerged in a stream during the summer trial. Day 0 samples are the epinecrotic bacterial community of the carcass prior to being placed into the stream.

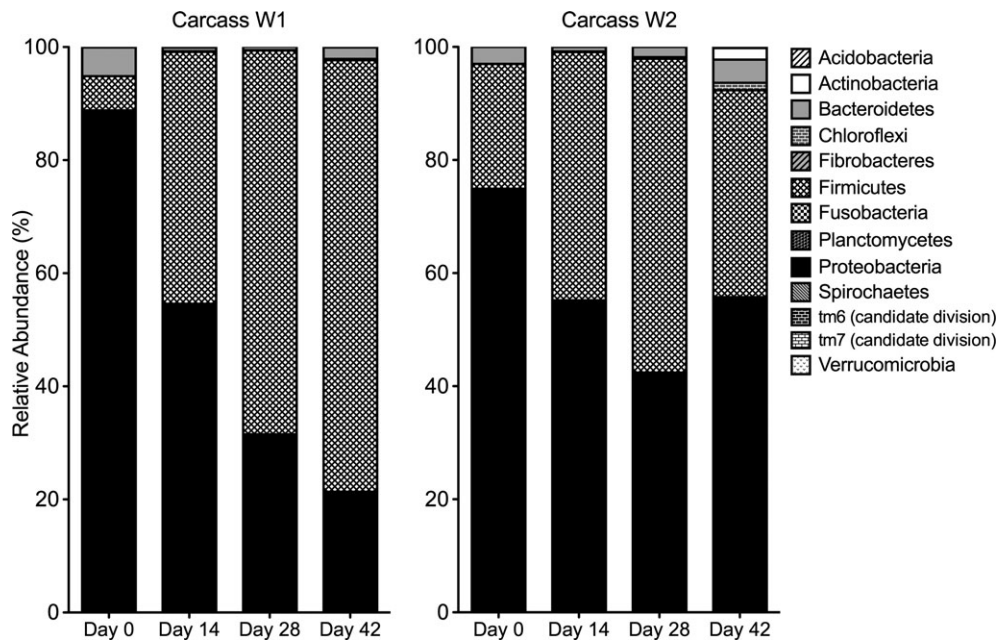


FIG. 3—Epinecrotic bacterial community taxa relative abundances change over decomposition of individual swine carcasses submerged in a stream for approximately 42 days during the winter trial. Day 0 samples are the epinecrotic bacterial community of the carcass prior to being placed into the stream.

between seasons ($p = 0.006$). During the summer, there were significant shifts in the bacterial communities at the phyletic and generic taxonomic levels of resolution among dates ($p < 0.05$). However, for winter, significant shifts in the communities were only significant among sampling dates ($p = 0.001$) at the generic level of taxonomic resolution.

Discussion

The focus of this study was to explore the use of high-throughput metagenomic sequencing of entire bacterial

communities associated with decomposing vertebrate remains for further use and development in forensics. We successfully identified epinecrotic bacterial communities throughout the decomposition process of carrion placed in a freshwater stream during two seasons, providing preliminary support for the potential use of these communities in forensic investigations. There were significantly different bacterial communities between summer and winter seasons, and depending on the taxonomic resolution employed to characterize the communities, there was a significant shift in these communities throughout the decomposition process. The differences in communities between seasons may

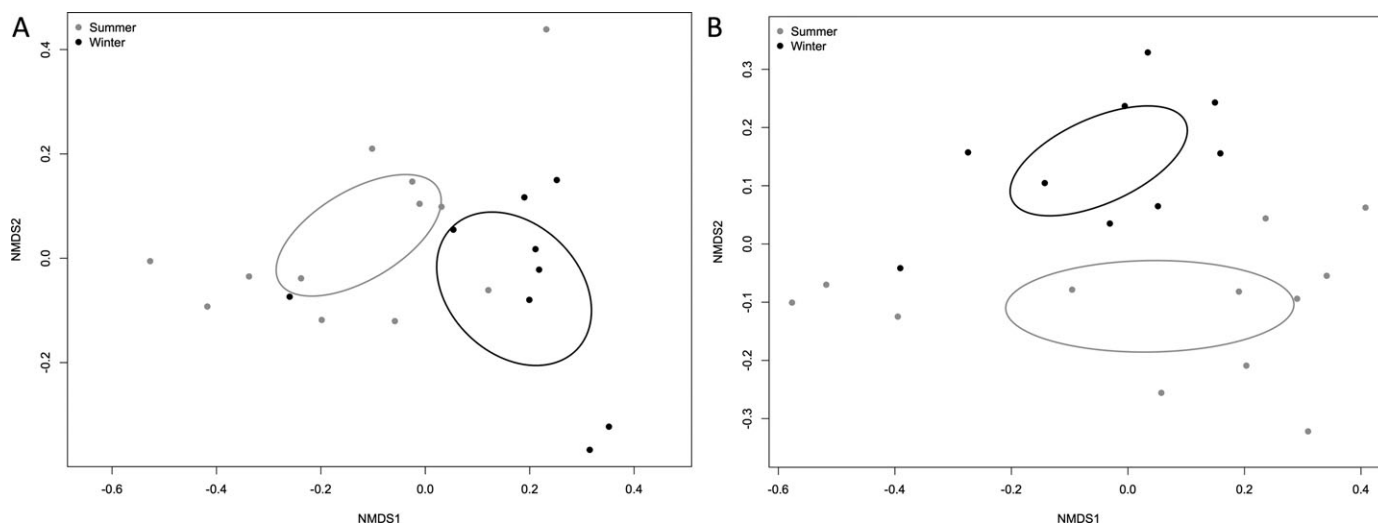


FIG. 4—Nonmetric multidimensional scaling analysis ordination of bacterial communities with significantly different community structures between seasonal trials (summer and winter) at the (A) phyletic (stress = 0.063, $R^2 = 0.98$) and (B) generic (stress = 0.152, $R^2 = 0.89$) taxonomic resolution. The ellipses indicate 95% standard error of each season.

be related to the abiotic conditions of the stream during those time periods, much like that reported for biofilms that form on inorganic substrates (69–71). Lower temperature and higher oxygen conditions during the winter trial may have selected for specific bacteria in the environment to initially colonize the carcasses, which were then replaced by bacteria that originally resided in the carcass (e.g., internal gut communities). It is well-known abiotic variables such as temperature, oxygen, and resource availability influence aquatic biofilm formation on nonliving surfaces (i.e., epilithic) in a way that affects community assembly (38,72–75). We suspect that this was the case in this study, shifting the bacterial taxa involved in the successional changes between winter and summer decomposition.

Further, carcass progression through decomposition during the summer followed an expected trajectory by taking an average of 22 days to decompose from submerged fresh to sunken remains. In the winter, the early floating decay stage was the last decomposition stage recorded prior to crayfish-scavenging events and vandalism, which prevented further characterization of the decomposition progress. However, winter carrion decomposition was approximately 2.3 times slower than the summer, requiring an average of 21 days until the early floating decay stage was reached, while in the summer it took an average of nine days. Based on these observations, the role of invertebrate scavenging on decomposing remains should be considered as a possible influence on evidence interpretation during investigations.

While there was intracarcass variation among carcasses, there were consistent bacterial community compositions for both summer and winter seasons; there was an inverse pattern of bacterial communities shifting from Proteobacteria as the predominant phyla in the early weeks of decomposition (submerged fresh to early floating decay) to the communities being dominated by Firmicutes in the latter portions of decomposition. In the summer trial, the predominant genera detected within Proteobacteria are known to be associated with vertebrate mammals and aquatic habitats. These genera include the following: *Pseudomonas* is common to aquatic systems (76) and has been found during the lipolysis process in adipocere formation (77); *Psychrobacter* has been reported from the nasal passageways of healthy swine (78); and *Ewingella* has been associated with meat spoilage (79,80). *Klebsiella*, which was detected during early decomposition in

the summer, has been isolated from swine carcasses (81) and is commonly detected in aquatic habitats such as lakes (82). Other predominate genera detected during the decomposition process are typical of polluted aquatic systems: *Citrobacter* are coliforms that likely entered the watershed through septic systems (83); *Zoogloea* has been cultivated from vegetation, specifically filamentous algae, and is commonly reported from organically polluted watersheds (84); and increases in *Dechloromonas* may result from changes in anthropogenic land use (wastewater treatment plant effluent), as previously reported in suburban and urban rivers (85).

In the winter trial, there were several genera of notable interest. For instance, *Desulfosporomusa* are sulfate-reducing bacteria that have been detected in sediments of an oligotrophic lake (86). *Clostridium* has been detected in numerous environments (87,88) and is a common part of the human gut flora, which can be pathogenic (89). *Enterococcus* is a genera used worldwide as a fecal coliform indicator, and can be cultured from fresh lake water after 30 days at 4°C (90). Furthermore, in aquatic habitats with vegetation, there are significantly higher *Enterococcus* densities detected in the sediments than in habitats without vegetation (91). *Proteocatella* has been detected in penguin guano (92), in the canine oral microbiome (93), and in dairy wastewater storage ponds (94). *Veillonellaceae*, which is closely related to *Clostridium*, has been found in a variety of habitats ranging from the human oral microbiome (95) to crude oil (96). Thus, the bacterial communities reported in this study represent taxa that can be commonly found in aquatic habitats and those associated with common vertebrate species used in human food production. This is not surprising and may reflect the environmental circumstances of decomposition within the watershed. However, we do report significant changes in the overall communities that reflect time since submersion, successional descriptions that have potential use as metrics for estimating PMSIs in future studies or case work.

Numerous environmental factors, such as physical and chemical variables (e.g., temperature and oxygen content) and biological mechanisms (e.g., bacteria, fungi, and detritivores), play a substantial role in the decomposition of human remains in aquatic systems, thus directly influencing estimates of a PMSI range (10,11). Microbial communities vary by geographic

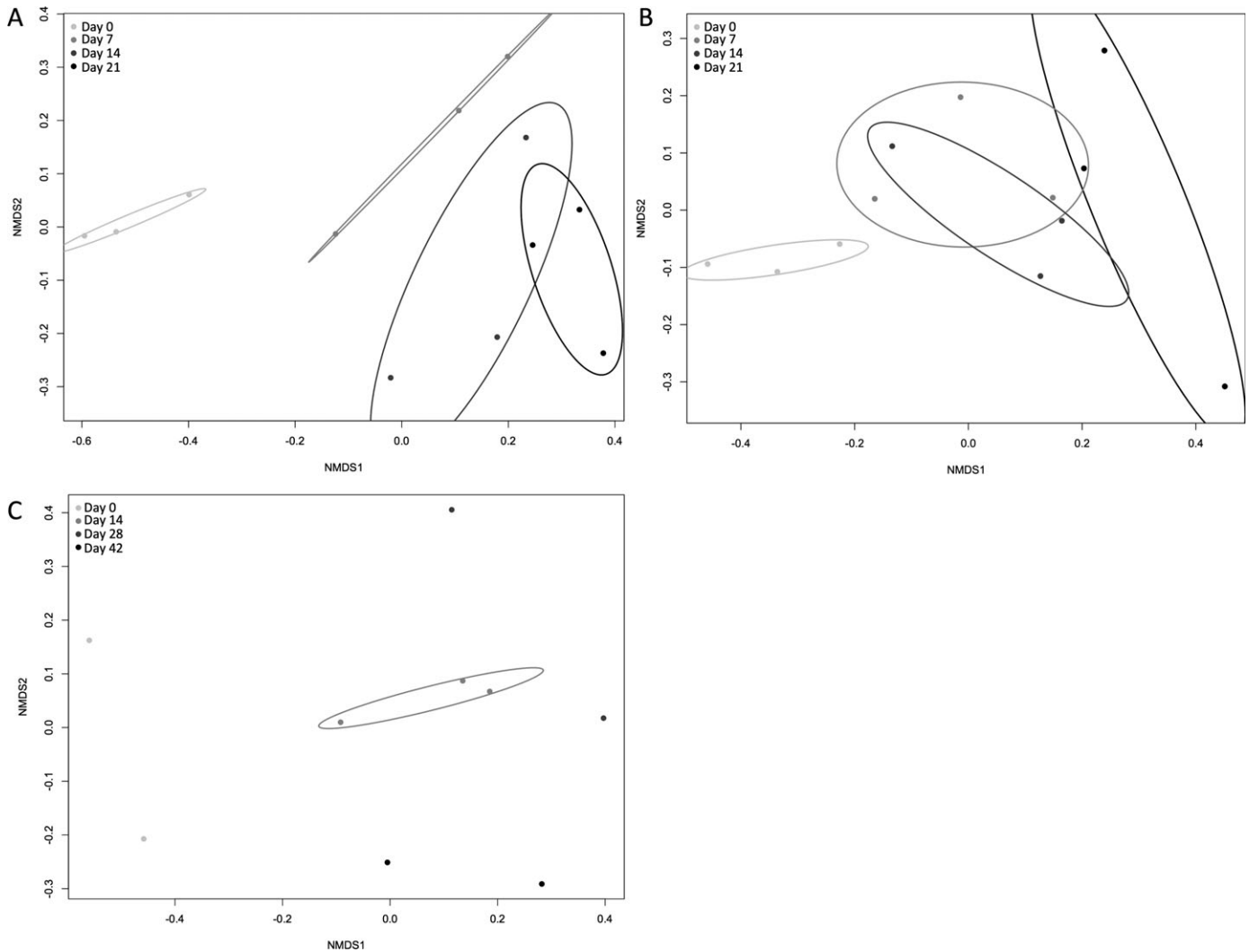


FIG. 5—Nonmetric multidimensional scaling analysis ordination of bacterial communities among sampling days during each seasonal trial (summer and winter). (A) There were significantly different bacterial communities at both the (A) generic (stress = 0.089, $R^2 = 0.96$) and (B) phyletic (stress = 0.045, $R^2 = 0.99$) taxonomic resolution during the summer trial. There were only sufficient data points at the (C) generic level (stress = 0.0111, $R^2 = 0.92$) to visualize significant differences among sampling days during the winter trial. The ellipses, when present, indicate 95% standard error of each sampling day.

TABLE 2—PERMANOVA results testing bacterial community structure responses at the phyletic and generic taxonomic levels between seasons, and among carcass decomposition days within each season (summer and winter). Significant effects are indicated by an asterisk.

Taxonomic Level	Factor	Source	d.f.	SS	MS	F	P
Phylum	Season	Season	1	0.698	0.698	5.53	0.006*
		Residuals	19	2.40	0.126		
		Total	20	3.10			
	Summer	Days	3	0.880	0.294	4.22	0.006*
		Residuals	8	0.557	0.070		
		Total	11	1.437			
	Winter	Days	3	0.419	0.140	1.29	0.273
		Residuals	5	0.543	0.109		
		Total	8	0.962			
Genera	Season	Season	1	0.942	0.942	2.77	0.002*
		Residuals	19	6.46	0.340		
		Total	20	7.40			
	Summer	Days	3	2.11	0.703	3.01	0.003*
		Residuals	8	1.87	0.233		
		Total	11	4.00			
	Winter	Days	3	1.35	0.449	1.99	0.001
		Residuals	5	1.13	0.226		
		Total	8	2.48			

region, season, and impacts of land use and land cover, thus possibly affecting how they could be used in forensic investigations, similar to the variability of using insects during investigations. And although the role of bacteria and fungi in both standing and flowing aquatic systems has been clearly elucidated (97–100), we are not aware of any study that has defined the role of stream or pond bacterial communities, the importance of specific taxa, or the succession of different microorganisms associated with the decomposition of carrion or human remains. Additional studies are warranted to more fully test and develop this potential metagenomic approach in forensic investigations.

While this study is novel and advances the knowledge of vertebrate remains decomposition in an aquatic habitat and the potential applications in forensics, there are limitations to the reference data set provided in this paper. Although conducted in a single stream (due to logistical constraints, carcasses were separated by at least one riffle/pool sequence in a single stream to minimize upstream influence of carcasses) and limited sample size, to the best of our knowledge this study documents the first high-throughput metagenomic approach identifying bacterial communities throughout carrion vertebrate decomposition in a freshwater habitat. Additionally, the use of pyrosequencing for

carion decomposition has its own limitations, as previously discussed by Pechal et al. (4). The current bacterial taxa information provided would be an excellent beginning for future research directions to more comprehensively explore the application of aquatic epinecrotic microbial communities in the forensic sciences.

Acknowledgments

JRW would like to acknowledge the support of the Mr. Mark Kreidler, Penn State University Swine Research Facility for providing the stillborn piglets for this work. Two anonymous reviewers provided constructive critical input that improved this paper.

References

- National Research Council. Strengthening forensic science in the United States: a path forward. Washington, DC: The National Academic Press, 2009.
- Tomberlin JK, Benbow ME, Tarone AM, Mohr R. Basic research in evolution and ecology enhances forensics. *Trends Ecol Evol* 2011;26:53–5.
- Hyde ER, Haarmann DP, Lynne AM, Bucheli SR, Petrosino JF. The living dead: bacterial community structure of a cadaver at the onset and end of the bloat stage of decomposition. *PLoS ONE* 2013;8:e77733.
- Pechal JL, Crippen TL, Benbow ME, Tarone AM, Dowd S, Tomberlin JK. The potential use of bacterial community succession in forensics as described by high throughput metagenomic sequencing. *Int J Legal Med* 2014;128:193–205.
- Metcalf JL, Parfrey LW, Gonzalez A, Lauber CL, Knights D, Ackermann G, et al. A microbial clock provides an accurate estimate of the postmortem interval in a mouse model system. *eLife* 2013;2:e01104.
- Rothberg JM, Leamon JH. The development and impact of 454 sequencing. *Nat Biotechnol* 2008;26:1117–24.
- Hudson ME. Sequencing breakthroughs for genomic ecology and evolutionary biology. *Mol Ecol Resour* 2008;8:3–17.
- The Human Microbiome Jumpstart Reference Strains Consortium. A catalog of reference genomes from the human microbiome. *Science* 2010;328:994–9.
- Zimmerman KA, Wallace JR. The potential to determine a postmortem submersion interval based on algal/diatom diversity on decomposing mammalian carcasses in brackish ponds in Delaware. *J Forensic Sci* 2008;53:935–41.
- Merritt RW, Wallace JR. The role of aquatic insects in forensic investigations. In: Byrd JH, Castner JL, editors. *Forensic entomology: the utility of arthropods in legal investigations*, 2nd edn. Washington, DC: CRC Press, 2010;177–222.
- Haefner JN, Wallace JR, Merritt RW. Pig decomposition in lotic aquatic systems: the potential use of algal growth in establishing a post-mortem submersion interval (PMSI). *J Forensic Sci* 2004;49:1–7.
- Costerton WJ, Lewandowski Z, Caldwell DE, Korber DR, Lappin-Scott HM. Microbial biofilms. *Annu Rev Microbiol* 1995;49:711–45.
- Giller PS, Malmqvist B. The biology of streams and rivers. New York, NY: Oxford University Press Inc., 1998.
- Battin TJ, Kaplan LA, Newbold JD, Hansen CM. Contributions of microbial biofilms to ecosystem processes in stream mesocosms. *Nature* 2003;426:439–42.
- Battin TJ, Sloan WT, Kjelleberg S, Daims H, Head IM, Curtis TP, et al. Microbial landscapes: new paths to biofilm research. *Nat Rev Microbiol* 2007;5:76–81.
- Costerton JW, Stewart PS, Greenberg EP. Bacterial biofilms: a common cause of persistent infections. *Science* 1999;284:1318–22.
- Freeman C, Lock MA. The biofilm polysaccharide matrix: a buffer against changing organic substrate supply? *Limnol Oceanogr* 1995;40:273–8.
- Sutherland IW. The biofilm matrix – an immobilized but dynamic microbial environment. *Trends Microbiol* 2001;9:222–7.
- Augsburger C, Karwautz C, Muflmann M, Daims H, Battin TJ. Drivers of bacterial colonization patterns in stream biofilms. *FEMS Microbiol Ecol* 2010;72:47–57.
- Freeman C, Chapman PJ, Gilman K, Lock MA, Reynolds B, Wheeler HS. Ion exchange mechanisms and the entrapment of nutrients by river biofilms. *Hydrobiologia* 1995;297:61–5.
- Sinsabaugh RL, Rept D, Weiland T, Golladay SW, Linkins AE. Exoenzyme accumulation in epilithic biofilms. *Hydrobiologia* 1991;222:29–37.
- Romaní AM, Fischer H, Mille-lindblom C, Tranvik LJ, Tranvik J, Romaní AM. Interactions of bacteria and fungi on decomposing litter: differential extracellular enzyme activities. *Ecol* 2006;87:2559–69.
- Tank JL, Dodds WK. Nutrient limitation of epilithic and epixylic biofilms in ten North American streams. *Freshw Biol* 2003;48:1031–49.
- Sinsabaugh RL, Golladay SW, Linkins AE. Comparison of epilithic and epixylic biofilm development in a boreal river. *Freshw Biol* 1991;25:179–87.
- Hladysz S, Cook RA, Petrie R, Nielsen DL. Influence of substratum on the variability of benthic biofilm stable isotope signatures: implications for energy flow to a primary consumer. *Hydrobiologia* 2011;664:135–46.
- Casamatta DA, Verb RG. Algal colonization of submerged carcasses in a mid-order woodland stream. *J Forensic Sci* 2000;45:1280–5.
- Dickson GC, Poulter RTM, Maas EW, Probert PK, Kieser JA. Marine bacterial succession as a potential indicator of postmortem submersion interval. *Forensic Sci Int* 2011;209:1–10.
- Claeson SM, Li JL, Compton JE, Bisson PA. Response of nutrients, biofilm, and benthic insects to salmon carcass addition. *Can J Fish Aquat Sci* 2006;63:1230–41.
- Parmenter RR, Lamarra VA. Nutrient cycling in a freshwater marsh: the decomposition of fish and waterfowl carrion. *Limnol Oceanogr* 1991;36:976–87.
- Tomberlin JK, Adler PH. Seasonal colonization and decomposition of rat carrion in water and on land in an open field in South Carolina. *J Med Entomol* 1998;35:704–9.
- Pohlon E, Marxsen JR, Kosel K. Pioneering bacterial and algal communities and potential extracellular enzyme activities of stream biofilms. *FEMS Microbiol Ecol* 2010;71:364–73.
- Ward DM, Ferris MJ, Nold SC, Bateson MM. A natural view of microbial biodiversity within hot spring cyanobacterial mat communities. *Microbiol Mol Biol R* 1998;62:1353–70.
- Carter DO, Tibbett M. Microbial decomposition of skeletal muscle tissue (*Ovis aries*) in a sandy loam soil at different temperatures. *Soil Biol Biochem* 2006;38:1139–45.
- Schimel JP, Gullledge JM, Clein-Curley JS, Lindstrom JE, Braddock JF. Moisture effects on microbial activity and community structure in decomposing birch litter in the Alaskan taiga. *Soil Biol Biochem* 1999;31:831–8.
- Ibekwe AM, Kennedy AC, Frohne PS, Papiernik SK, Yang CH, Crowley DE. Microbial diversity along a transect of agronomic zones. *FEMS Microbiol Ecol* 2002;39:183–91.
- Kuske CR, Ticknor LO, Miller ME, Dunbar JM, Davis JA, Barns SM, et al. Comparison of soil bacterial communities in rhizospheres of three plant species and the interspaces in an arid grassland. *Appl Environ Microbiol* 2002;68:1854–63.
- Haslam TCF, Tibbett M. Soils of contrasting pH affect the decomposition of buried mammalian (*Ovis aries*) skeletal muscle tissue. *J Forensic Sci* 2009;54:900–4.
- Besemer K, Peter H, Logue JB, Langenheder S, Lindstrom S, Tranvik LJ, et al. Unraveling assembly of stream biofilm communities. *ISME J* 2012;6(8):1459–68.
- Haskell N, McShaffrey D, Hawley D, Williams R, Pless J. Use of aquatic insects in determining submersion interval. *J Forensic Sci* 1989;34:622–32.
- Hawley D, Haskell N, McShaffrey D, Williams R, Pless J. Identification of a red “fiber”: chironomid larvae. *J Forensic Sci* 1989;34:617–21.
- Keiper JB, Chapman EG, Foote BA. Midge larvae (Diptera: Chironomidae) as indicators of postmortem submersion interval of carcasses in a woodland stream: a preliminary report. *J Forensic Sci* 1997;42:1074–9.
- Siver P, Lord W, McCarthy D. Forensic limnology: the use of freshwater algal community ecology to link suspects to an aquatic crime scene in southern New England. *J Forensic Sci* 1994;39:847.
- Wallace JR, Merritt RW, Kimbirauskas R, Benbow ME, McIntosh M. Caddisflies assist with homicide case: determining a postmortem submersion interval using aquatic insects. *J Forensic Sci* 2008;53:219–21.
- Hobischak N, Anderson G. Time of submergence using aquatic invertebrate succession and decomposition changes. *J Forensic Sci* 2002;47:142–51.
- Keiper JB, Casamatta DA. Benthic organisms as forensic indicators. *J North Am Benthol Soc* 2001;20:311–24.

46. Schoenly KG, Haskell NH, Hall RD, Gbur JR. Comparative performance and complementarity of four sampling methods and arthropod preference tests from human and porcine remains at the Forensic Anthropology Center in Knoxville, Tennessee. *J Med Entomol* 2007;44:881–94.
47. Giotis ES, Loeffler A, Knight-Jones T, Lloyd DH. Development of a skin colonization model in gnotobiotic piglets for the study of the microbial ecology of methicillin-resistant *Staphylococcus aureus* ST398. *J Appl Microbiol* 2012;113:992–1000.
48. Vodica P, Smetana K, Dvorankova B, Emerick T, Xu YZ, Ourednik J, et al. The miniature pig as an animal model in biomedical research. *Ann N Y Acad Sci* 2005;1049:161–71.
49. Dick IP, Scott RC. Pig ear skin as an *in-vitro* model for human skin permeability. *J Pharm Pharmacol* 1992;44:640–5.
50. Allaker RP. Prevention of exudative dermatitis in gnotobiotic piglets by bacterial interference. *Vet Rec* 1988;123:597–8.
51. Miller DN, Bryant JE, Madsen EL, Ghiorse WC. Evaluation and optimization of DNA extraction and purification procedures for soil and sediment samples. *Appl Environ Microbiol* 1999;65:4715–24.
52. Zhou J, Bruns MA, Tiedje JM. DNA recovery from soils of diverse composition. *Appl Environ Microbiol* 1996;62:316–22.
53. Lear G, Dong Y, Lewis G. Comparison of methods for the extraction of DNA from stream epilithic biofilms. *Anton Leeuw Int J G* 2010;98:567–71.
54. Dowd SE, Wolcott RD, Sun Y, McKeen T, Smith E, Rhoads D. Polymicrobial nature of chronic diabetic foot ulcer biofilm infections determined using bacterial tag encoded FLX amplicon pyrosequencing (bTEFAP). *PLoS ONE* 2008;3:e3326.
55. Sen R, Ishak HD, Estrada D, Dowd SE, Hong E, Mueller UG. Generalized antifungal activity and 454-screening of *Pseudonocardia* and *Amycolatopsis* bacteria in nests of fungus-growing ants. *Proc Natl Acad Sci USA* 2009;106:17805–10.
56. Handl S, Dowd SE, Garcia-Mazcorro JF, Steiner JM, Suchodolski JS. Massive parallel 16S rRNA gene pyrosequencing reveals highly diverse fecal bacterial and fungal communities in healthy dogs and cats. *FEMS Microbiol Ecol* 2011;76:301–10.
57. Andreotti R, Perez de Leon AA, Dowd SE, Guerrero FD, Bendele KG, Scoles GA. Assessment of bacterial diversity in the cattle tick *Rhipicephalus (Boophilus) microplus* through tag-encoded pyrosequencing. *BMC Microbiol* 2011;11:1–6.
58. La Duc MT, Vaishampayan P, Nilsson HR, Torok T, Venkateswaran K. Pyrosequencing-derived bacterial, archaeal, and fungal diversity of spacecraft hardware destined for mars. *Appl Environ Microbiol* 2012;78:5912–22.
59. Dowd SE, Callaway TR, Wolcott RD, Sun Y, McKeen T, Hagevoort RG, et al. Evaluation of the bacterial diversity in the feces of cattle using 16S rDNA bacterial tag-encoded FLX amplicon pyrosequencing (bTEFAP). *BMC Microbiol* 2008;8:125.
60. Dowd SE, Sun Y, Secor PR, Rhoads DD, Wolcott BM, James GA, et al. Survey of bacterial diversity in chronic wounds using pyrosequencing, DDGE, and full ribosome shotgun sequencing. *BMC Microbiol* 2008;8:43.
61. Edgar RC. Search and clustering orders of magnitude faster than BLAST. *Bioinform* 2010;26:2460–1.
62. Capone KA, Dowd SE, Stamatas GN, Nikolovski J. Diversity of the human skin microbiome early in life. *J Invest Dermatol* 2011;131:2026–32.
63. Dowd S, Hanson JD, Rees E, Wolcott R, Zischau A, Sun Y, et al. Research survey of fungi and yeast in polymicrobial infections in chronic wounds. *J Wound Care* 2011;20:40–7.
64. DeSantis TZ, Hugenholtz P, Larsen N, Rojas M, Brodie EL, Keller K, et al. Greengenes, a chimera-checked 16S rRNA gene database and workbench compatible with ARB. *Appl Environ Microbiol* 2006;72:5069–72.
65. Simpson EH. Measurement of diversity. *Nature* 1949;163:688.
66. R Development Core Team. R: a language and environment for statistical computing. Vienna, Austria, 2010;ISBN 3-900051-07-0; <http://www.R-project.org> (accessed December 4, 2013).
67. McCune B, Grace J. Analysis of ecological communities. Gleneden Beach, OR: MjM, 2002.
68. Anderson MJ. A new method for non-parametric multivariate analysis of variance. *Austral Ecol* 2001;26:32–46.
69. Wimpenny J. An overview of biofilms as functional communities. In: Allison DG, Gilbert P, Lappin-Scott HM, Wilson M, editors. Proceedings of the fifty-ninth symposium of the society for general microbiology; 2000 Sept 12–15; Exeter, U.K. Cambridge, U.K.: Cambridge University Press, 2000;1–24.
70. Roman AM, Giorgi A, Acuna V, Sabater S. The influence of substrate type and nutrient supply on biofilm organic matter utilization in streams. *Limnol Oceanogr* 2004;49:1713–21.
71. Roman AM, Sabater S. Effect of primary producers on the heterotrophic metabolism of a stream biofilm. *Freshw Biol* 1999;41:729–36.
72. Besemer K, Singer G, Hodl I, Battin TJ. Bacterial community composition of stream biofilms in spatially variable-flow environments. *Appl Environ Microbiol* 2009;75:7189–95.
73. Pekkonen M, Ketola T, Laakso JT. Resource availability and competition shape the evolution of survival and growth ability in a bacterial community. *PLoS ONE* 2013;8:e76471.
74. Lear G, Anderson MJ, Smith JP, Boxen K, Lewis GD. Spatial and temporal heterogeneity of the bacterial communities in stream epilithic biofilms. *FEMS Microbiol Ecol* 2008;65:463–73.
75. Lyautey E, Jackson CR, Cayrou J, Rols J-L, Garabétian F. Bacterial community succession in natural river biofilm assemblages. *Microbiol Ecol* 2005;50:589–601.
76. Römmling U, Wingender J, Müller H, Tümmler B. A major *Pseudomonas aeruginosa* clone common to patients and aquatic habitats. *Appl Environ Microbiol* 1994;60:1734–8.
77. Pfeiffer S, Milne S, Stevenson R. The natural decomposition of adipocere. *J Forensic Sci* 1998;43:368–70.
78. Weese JS, Slifert M, Jalali M, Friendship R. Evaluation of the nasal microbiota in slaughter-age pigs and the impact on nasal methicillin-resistant *Staphylococcus aureus* (MRSA) carriage. *BMC Vet Res* 2014;10:69.
79. Carter M, Chengappa M. Enterobacteria. In: Carter G, Cole J, editors. Diagnostic procedures in bacteriology and mycology, 5th edn. San Diego, CA: Academic Press Inc, 1990;107–28.
80. Lyon W, Milliet J. Microbial flora associated with Louisiana processed frozen and fresh nutria (*Myocastor coypus*) carcasses. *J Food Sci* 2000;65:1041–5.
81. Zou LK, Wang HN, Zeng B, Zhang AY, Li JN, Li XT, et al. Phenotypic and genotypic characterization of β -lactam resistance in *Klebsiella pneumoniae* isolated from swine. *Vet Microbiol* 2011;149:139–46.
82. Campbell LM, Michaels G, Klein RD, Roth IL. Isolation of *Klebsiella pneumoniae* from lake water. *Can J Microbiol* 1976;22:1762–7.
83. Janakiraman A, Leff LG. Comparison of survival of different species of bacteria in freshwater microcosms. *J Freshw Ecol* 1999;14:233–40.
84. Williams TM, Unz RF. Environmental distribution of *Zoogloea* strains. *Water Res* 1983;17:779–87.
85. Drury B, Rosi-Marshall E, Kelly JJ. Wastewater treatment effluent reduces the abundance and diversity of benthic bacterial communities in urban and suburban rivers. *Appl Environ Microbiol* 2013;79:1897–905.
86. Sass H, Overmann J, Rütters H, Babenzien H-D, Cypionka H. *Desulfosporomusa polytropica* gen. nov., sp. nov., a novel sulfate-reducing bacterium from sediments of an oligotrophic lake. *Arch Microbiol* 2004;182:204–11.
87. Lalitha KV, Gopakumar K. Distribution and ecology of *Clostridium botulinum* in fish and aquatic environments of a tropical region. *Food Microbiol* 2000;17:535–41.
88. Byamukama D, Mach RL, Kansiime F, Manafi M, Farnleitner AH. Discrimination efficacy of fecal pollution detection in different aquatic habitats of a high-altitude tropical country, using presumptive coliforms, *Escherichia coli*, and *Clostridium perfringens* spores. *Appl Environ Microbiol* 2005;71:65–71.
89. van der Waaij D. The ecology of the human intestine and its consequences for overgrowth by pathogens such as *Clostridium difficile*. *Annu Rev Microbiol* 1989;43:69–87.
90. Lleó MDM, Bonato B, Benedetti D, Canepari P. Survival of enterococcal species in aquatic environments. *FEMS Microbiol Ecol* 2005;54:189–96.
91. Badgley BD, Thomas FI, Harwood VJ. The effects of submerged aquatic vegetation on the persistence of environmental populations of *Enterococcus* spp. *Environ Microbiol* 2010;12:1271–81.
92. Pikuta EV, Hoover RB, Marsic D, Whitman WB, Lupa B, Tang J, et al. *Proteocatella sphenisci* gen. nov., sp. nov., a psychrotolerant, spore-forming anaerobe isolated from penguin guano. *Int J Syst Evol Microbiol* 2009;59:2302–7.
93. Dewhirst FE, Klein EA, Thompson EC, Blanton JM, Chen T, Milella L, et al. The canine oral microbiome. *PLoS ONE* 2012;7:e36067.

94. Dungan RS, Leytem AB. The characterization of microorganisms in dairy wastewater storage ponds. *J Environ Qual* 2013;42:1583–8.
95. Dewhirst FE, Chen T, Izard J, Paster BJ, Tanner ACR, Yu WH, et al. The human oral microbiome. *J Bacteriol* 2010;192:5002–17.
96. Albokari M, Mashhour I, Alshehri M, Boothman C, Al-Enezi M. Characterization of microbial communities in heavy crude oil from Saudi Arabia. *Ann Microbiol* 2015;65(1):95–104.
97. Baldy V, Gessner MO, Chauvet E. Bacteria, fungi and the breakdown of leaf litter in a large river. *Oikos* 1995;74:93–102.
98. Bengtsson G. Interactions between fungi bacteria and beech leaves in a stream microcosm. *Oecologia* 1992;89:542–9.
99. Kaushik N, Hynes H. The fate of dead leaves that fall into streams. *Arch Hydrobiol* 1971;68:465–515.
100. Suberkropp K, Klug MJ. Fungi and bacteria associated with leaves during processing in a woodland stream. *Ecol* 1976;57:707–19.

Additional information and reprint requests:

M. Eric Benbow, Ph.D.

Department of Entomology and Department of Osteopathic Medical
Specialties

Michigan State University

243 Natural Science Building

East Lansing, MI 48824

E-mail: benbow@msu.edu

Supporting Information

Additional Supporting Information may be found in the online version of this article:

Table S1 A visual assessment of the decomposition process for replicate swine carcasses (n=3) placed in a temperate head-water stream during two seasons (summer and winter) was determined according to Zimmerman and Wallace (9).

Table S2 Mean \pm (standard deviation) bacterial taxa relative abundance as characterized by pyrosequencing and universal 16s rRNA primers over decomposition at weekly sampling intervals during the summer season.

Table S3 Mean \pm (standard deviation) bacterial taxa relative abundance as characterized by pyrosequencing and universal 16s rRNA primers over decomposition at weekly sampling intervals during the winter season.